

Mapping soil organic carbon under erosion processes using remote sensing

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Abstract

This study aimed to map soil organic carbon under erosion processes on an arable field in the Republic of Bashkortostan (Russia). To estimate the spatial distribution of organic carbon in the Haplic Chernozem topsoil, we applied Sentinel-2A satellite data and the linear regression method. We used 13 satellite bands and 15 calculated spectral indices for regression modelling. A regression model with an average prediction level has been created ($R^2 = 0.58$, RMSE = 0.56, RPD = 1.61). Based on the regression model, cartographic materials for organic carbon content have been created. Water flows and erosion processes were determined using the calculated Flow Accumulation model. The relationship between organic carbon, biological activity, and erosion conditions is shown. The ¹³C-NMR spectroscopy method was used to estimate the content and nature of humic substances of different soil samples. Based on the ¹³C-NMR analysis, a correlation was established with the spectral reflectivity of eroded and non-eroded soils. It was revealed that the effect of soil organic carbon on spectral reflectivity depends not only on the quantity but also on the quality of humic substances and soil formation conditions.

Keywords: Soil organic carbon, remote sensing, sentinel, erosion, humic acids, ¹³C-NMR

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Introduction

Water erosion is one of the most dangerous processes due to the transformation of climatic conditions and anthropogenic impacts. Erosion of agricultural lands and subsequent fertility decline is one of the reasons for the abandonment of lands (BAUDE, M. *et al.* 2019; SULEYMANOV R. *et al.* 2020a). This is especially true for the southern regions of Russia, where degradation processes proceed at an accelerated pace (KATTISOV, V.M. 2017). The Republic of Bashkortostan is located in the southern part

of the Ural Mountains and characterized by actively occurring processes of water and wind erosion (SOBOL, N.V. *et al.* 2015; GABBASOVA, I.M. *et al.* 2016; SULEYMANOV, R. *et al.* 2019). The size of erosion rates can be judged by the following data: the land fund of the republic is 142,970 km², agricultural land occupies 73,430 km² (51.37%), 36,000 km² are erosion-hazardous (25.18%), 33,000 km² are exposed to water erosion (23.08%), wind erosion – 10,500 km² (7.35%), the joint action of water and wind erosion – 120 km² (0.08%) (GABBASOVA, I.M. *et al.* 2016; SULEYMANOV, R. *et al.* 2020b).

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Remote sensing (RS) is a useful tool in soil researches (MULDER, V.L. *et al.* 2011; SAVIN, I.YU. *et al.* 2019). Multi- and hyperspectral images from unmanned aerial vehicles, aircraft, and space satellites are used for different scientific tasks. The active use of satellite data in last years is facilitated by improved spatial resolution, a large data set (multi-year image archives), a short interval, free access to satellite images (Sentinel, Landsat, and others) (PRUDNIKOVA, E.YU. and SAVIN, I.YU. 2015; ANGELOPOULOU, T. *et al.* 2019). RS methods are more cost-effective and allow to cover large areas.

The integration of RS and GIS is a valuable tool for research, digital mapping, and modeling of erosion processes (LEH, M. *et al.* 2013; GUO, B. *et al.* 2018; NAMPAAK, H. *et al.* 2018; SULEYMANOV, A.R. 2019; YANG, X. *et al.* 2020). Study soil erosion based on RS and GIS methods are currently being actively studied throughout the world (DESPRATS, J.F. *et al.* 2013; WANG, L. *et al.* 2013; PANAGOS, P. *et al.* 2015; YERMOLAEV, O.P. 2017; GOLOSOV, V. *et al.* 2018; FRANKL, A. *et al.* 2018; SEPURU, T.K. and DUBE, T. 2018; PHINZI, K. and NGETAR, N.S. 2019; MAGLIULO, P. *et al.* 2020). Erosion processes are directly related to soil organic carbon (SOC) content. In a review article by ANGELOPOULOU, T. *et al.* (2019), about evaluating SOC based on RS data, the authors conclude that recent advances in machine learning can help improve the overall accuracy and reliability of models. Thus, many studies confirm the successful use of satellite data in the study of transformation processes and automated mapping of SOC content (MOUAZEN, A.M. *et al.* 2007; BARTHOLOMEUS, H. *et al.* 2008; DUBE, T. *et al.* 2018; GHOLIZADEH, A. *et al.* 2018; BHUNIA, G.S. *et al.* 2019; CASTALDI, F. *et al.* 2019; CHEN, D. *et al.* 2019; DOU, X. *et al.* 2019; VAUDOUR, E. *et al.* 2019).

According to some studies (BEN-DOR, E. *et al.* 1997; VISCARRA ROSSEL, R.A. *et al.* 2006b; NOCITA, M. *et al.* 2015; CASTALDI, F. *et al.* 2018, 2019), in order to model and predict the SOC content, recommended to use spectral characteristics located in the visible range at 450, 590, 664 nm, as well as the characteristics of the invisible range in the short-wave infrared SWIR range – between 1,600 and 1,900 nm and

about 2,100 and 2,300 nm. At satellite Sentinel-2A, the bands of the visible spectrum B2, B3, B4 (490, 560, and 665 nm, respectively) and the invisible range in the SWIR region are B11 and B12 (1,610 and 2,190 nm, respectively). Thus, as Sentinel-2A has a close correspondence of the spectral characteristics, it allows the use of data for the estimation and modelling of SOC content in topsoil.

Soil colour is one of the key indicators for digital SOC mapping using RS (MULDER, V.L. *et al.* 2011). The qualitative and quantitative composition of humic acids in turn, affects the colour of the soil (VISCARRA ROSSEL, R.A. *et al.* 2006a). Thus, humic substances are an indicator of the spectral reflectivity of the soil, with which a remote evaluation is possible. At the same time, soil organic matter (SOM) plays an important role in maintaining a good soil structure and directly affects the nature of erosion processes. SOM is responsible for the content of organic substances, nutrients, the activity of microorganisms, moisture retention.

Nuclear Magnetic Resonance (NMR) spectroscopy is a highly accurate physical and chemical tool for determining the composition and structure of soil organic matter (QUIDEAU, S.A. *et al.* 2000; CHUKOV, S.N. *et al.* 2017, 2018; POLYAKOV, V. and ABAKUMOV, E.V. 2020). The ¹³C-NMR spectroscopy method allows to study the structural and compositional features of humic acid preparations of eroded soils, which will help to understand the ongoing processes of humification, degradation, and SOC transfer on eroded lands (SIMPSON, M.J. *et al.* 2008; ABAKUMOV, E.V. *et al.* 2013; RUMPEL, C. *et al.* 2014; CONTE, P. *et al.* 2017).

Currently, in conditions of active anthropogenic impact, it is necessary to conduct local studies to understand the processes of transformation and degradation of SOC. Thus, our work aims to map SOC content, estimate and study transport processes on an eroded agricultural field, using Sentinel-2A satellite data, ¹³C-NMR spectroscopy, and geomorphological methods. The developed methodology will allow us to simulate and evaluate the SOC content in the topsoil of agricultural land for monitoring and mapping.

Materials and methods

Site description

The study area is cropland (1,400 hectares) in long-term agricultural use located in the south of Russia, in the Zilair region of the Republic of Bashkortostan (Figure 1). The site is characterized by ploughing with a turnover of the soil layer at a depth of 10–15 cm. Wheat (*Triticum aestivum*) is predominantly growing on the plot. The cropland is located on gentle slopes of various exposures. According to geomorphometric analysis based on digital elevation model (DEM), the height of the study area varies from 460 m in the north-western part of the site to 377 m above sea level in the south-eastern part. The area mainly consists of slopes of up to 4°. The steepest sites are located in the southern, south-western, and northern parts of the site. The water erosion processes take place in the southern and northern parts of the territory. Wind erosion processes are also observed at the site.

The climate of the region is arid or slightly arid. The average annual air temperature is 1.4 °C, the average annual rainfall is 379 mm. According to the World Reference Base (WRB) for soil resources (IUSS Working Group WRB, 2014), the soil of the study site is characterized as Haplic Chernozems. The parent rocks are the eluvial-deluvial carbonate clays and heavy loams, as well as the eluvium of sandy schists.

Soil samples

The soil sampling work was carried out in October 2018 (49 full-profile sections and 5 pits). The soil samples were identified by satellite images to choose areas with different spectral reflectivity and erosion conditions. The exact coordinates of each soil point were identified using a global positioning system (GPS) with an accuracy of ± 3 m. The crop has already been harvested at this time. Samples for the analysis SOC content were taken from

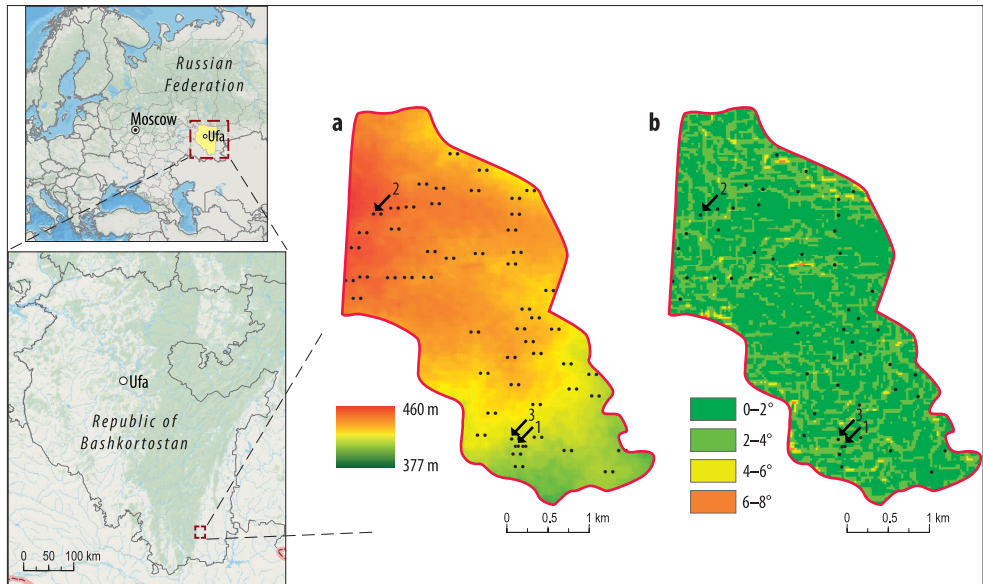


Fig. 1. Map of the study area (left) and spatial distribution of soil samples ($n = 54$). – a = elevation map in metres a.s.l.; b = slope map in degrees. Arrows and numbers (1–3) indicate samples for ^{13}C -NMR spectroscopy analysis.

the topsoil (0–10 cm). The carbon content was determined using the Tyurin method with colourimetric termination, according to Orlov and Grindel (ARINUSHKINA, E.V. 1970; SOKOLOV, A.V. 1975). The microbiological activity of soils, the basal respiration, using incubation chambers was determined by standard protocol (LAL, R. *et al.* 2001). Soil basal respiration is defined as the steady rate of respiration in soil, which originates from the mineralization of organic matter.

Remote sensing data

The Sentinel-2 satellite free-access dataset (Level-2A processing) was used for the study. The satellite data contains 13 spectral bands with a spatial resolution of 10 to 60 m. (Table 1). The cloud-free scenes from 02.10.2018 were selected for the study. This scene time is selected for work with bare soil and reduce vegetation impact. Then images went through the stages of atmospheric and radiometric correction using the module “Semi-Automatic Classification Plugin” in QGIS 3.6.0.

For more complex analysis, the most popular spectral indices for predicting soil attributes have been selected. The following indices based on the combination of Sentinel-2A satellite bands were calculated: Normalized Difference Vegetation Index (NDVI), Transformed Vegetation Index

(TVI), Enhanced Vegetation Index (EVI), Soil Adjusted Total Vegetation Index (SATVI), Soil Adjusted Vegetation Index (SAVI), Moisture Stress Index (MSI), Green Normalized Difference Vegetation Index (GNDVI), Green-Red Vegetation Index (GRVI), Land Surface Water Index (LSWI), Modified Soil Adjusted Vegetation Index (MSAVI), the Second Modified Soil Adjusted Vegetation Index (MSAVI2), Brightness Index (BI), the Second Brightness Index (BI2), Redness Index (RI), Color Index (CI). The index formulae and descriptions are presented in Table 2.

Vegetation indices are important predictors and are actively used in the modelling and mapping of soil properties. For example, BHUNIA, G.S. *et al.* (2019) successfully applied NDVI and BSI indices using a multivariate regression approach for SOC mapping. GHOLIZADEH, A. *et al.* (2018) showed that GNDVI and SATVI indices provided the strongest correlation with SOC on agricultural plots. Also, several studies conclude that vegetation indices are the most important variables in predicting soil properties (GOPP, N.V. *et al.* 2017; CHEN, D. *et al.* 2019; EMADI, M. *et al.* 2020).

¹³C-NMR spectroscopy

Soil samples for spectroscopy were selected according to the following parameters: a sample from a site of study area without water

Table 1. Sentinel-2A bands specifications

Band	Spectral range, nm	Spatial resolution, m	Spectral position, nm	Bandwidth, nm
B1	433–453	60	443	20
B2	458–523	10	490	65
B3	543–578	10	560	35
B4	650–680	10	665	30
B5	698–713	20	705	15
B6	733–748	20	740	15
B7	773–793	20	783	20
B8	785–900	10	842	115
B8a	855–875	20	865	20
B9	935–955	60	945	20
B10	1,360–1,390	60	1,380	30
B11	1,565–1,655	20	1,610	90
B12	2,100–2,280	20	2,190	180

Table 2. The calculated spectral indices

No.	Spectral index	Formula	Formula for Sentinel-2A	Details	References
1	NDVI	$\frac{\text{NIR} - \text{R}}{\text{NIR} + \text{R}}$	$\frac{\text{B8} - \text{B4}}{\text{B8} + \text{B4}}$	-	ROUSE, J.W., JR. <i>et al.</i> (1973)
2	GNDVI	$\frac{\text{NIR} - \text{G}}{\text{NIR} + \text{G}}$	$\frac{\text{B8} - \text{B3}}{\text{B8} + \text{B3}}$	-	GITTELSON, A.A. <i>et al.</i> (1996)
3	EVI	$A \left(\frac{\text{NIR} + \text{C1} \cdot \text{R} - \text{C2} \cdot \text{B} + \text{L}}{\text{NIR} + \text{R} + \text{L}} \right)$	$2.5 \left(\frac{\text{B8} - \text{B4}}{\text{B8} + 6 \cdot \text{B4} - 7.5 \cdot \text{B2} + 1} \right)$	A = 2.5 C1 = 6.0 C2 = 7.5 L = 1.0	HUETE, A. <i>et al.</i> (2002)
4	CI	$\frac{\text{R} - \text{G}}{\text{R} + \text{G}}$	$\frac{\text{B4} - \text{B3}}{\text{B4} + \text{B3}}$	-	POUGET, M. <i>et al.</i> (1990)
5	BI	$\frac{\sqrt{(\text{R} \cdot \text{R}) + (\text{G} \cdot \text{G})}}{2}$	$\frac{\sqrt{(\text{B4} \cdot \text{B4}) + (\text{B3} \cdot \text{B3})}}{2}$	-	ESCADAFAL, R. (1989)
6	BI2	$\frac{\sqrt{(\text{R} \cdot \text{R}) + (\text{G} \cdot \text{G}) + (\text{NIR} \cdot \text{NIR})}}{3}$	$\frac{\sqrt{(\text{B4} \cdot \text{B4}) + (\text{B3} \cdot \text{B3}) + (\text{B8} \cdot \text{B8})}}{3}$	-	ESCADAFAL, R. (1989)
7	TVI	$\frac{\text{NIR} - \text{R} + 0.5\text{I}/2 \cdot 100}{\text{NIR} + \text{R} + \text{L}}$	$\frac{\text{B8} - \text{B4} + 0.5\text{I}/2 \cdot 100}{\text{B8} + \text{B4} + \text{L}}$	-	NELLIS, M.D. and BRIGGS, J.M. (1992)
8	SAVI	$\frac{(\text{NIR} - \text{R}) \cdot (1 + \text{L})}{\text{NIR} - \text{R} + \text{L}}$	$\frac{(\text{B8} - \text{B4}) \cdot (1 + \text{L})}{\text{B8} - \text{B4} + \text{L}}$	L = 0.5	HUETE, A.R. (1988)
9	SATVI	$\frac{\text{SWIR1} - \text{R}}{\text{SWIR1} + \text{R} + \text{L}} \cdot (1 + \text{L}) - \text{SWIR2}/2$	$\frac{\text{B11} - \text{B4}}{\text{B11} + \text{B4} + \text{L}} \cdot (1 + \text{L}) - \text{B12}/2$	L = 1.0	MARSETT, R.C. <i>et al.</i> (2006)
10	RI	$\frac{\text{R} \cdot \text{R}}{\text{G} \cdot \text{G} \cdot \text{G}}$	$\frac{\text{B4} \cdot \text{B4}}{\text{B3} \cdot \text{B3} \cdot \text{B3}}$	-	POUGET, M. <i>et al.</i> (1990)
11	MSI	$\frac{\text{SWIR1}}{\text{NIR}}$	$\frac{\text{B11}}{\text{B8}}$	-	ROCK, B.N. <i>et al.</i> (1985)
12	LSWI	$\frac{\text{NIR} - \text{SWIR1}}{\text{NIR} + \text{SWIR1}}$	$\frac{\text{B8} - \text{B11}}{\text{B8} + \text{B11}}$	-	XIAO, X. <i>et al.</i> (2004)
13	GRVI	$\frac{\text{G} - \text{R}}{\text{G} + \text{R}}$	$\frac{\text{B3} - \text{B4}}{\text{B3} + \text{B4}}$	-	TUCKER, C.J. (1979)

Note: B, G, R, NIR, SWIR – reflection in blue, green, red, near infrared, short-wave infrared band, respectively.

erosion processes, with erosion processes and erosion sediment. Samples from the site without water erosion were taken at the top in the northern part of the site. Samples from the plot of erosion processes and sediment were taken at the southern part of the field (see *Figure 1*).

Humic acids (HAs) were extracted according to a published IHSS protocol (SWIFT, R.S. 1996). Solid-state CP/MAS ^{13}C -NMR spectra of HAs were obtained by Bruker Avance 500 NMR spectrometer. The repetition delay was 3 seconds. The number of scans was 6,500–29,000. Contact time is 0.2 μs .

Various molecular fragments were identified by CP/MAS ^{13}C -NMR spectroscopy (*Table 3*): carboxyl ($-\text{COOR}$); carbonyl ($-\text{C}=\text{O}$); CH_3- , CH_2- , CH -aliphatic; $-\text{C}-\text{OR}$ alcohols, esters, and carbohydrates; phenolic ($\text{Ar}-\text{OH}$); quinone ($\text{Ar}=\text{O}$); aromatic ($\text{Ar}-$), which indicates the great complexity of the structure of HAs and the poly-functional properties that cause their active participation in soil processes (LODYGIN, E.D. *et al.* 2014).

To standardize the quantitative characteristics of humic acid macromolecules, the following parameters were used: carbon ratio of aromatic structures to aliphatic structures, the decomposition rate of organic matter (C-Alkyl/ O-Alkyl), and humic acid hydrophobicity integral index ($\text{AL h,r} + \text{AR h,r}$, %).

Geomorphometric analysis

The geomorphometric analysis of the territory was carried out using the QGIS and SAGA GIS based on a digital elevation model (DEM) with a resolution of 30 m – NASA's

Shuttle Radar Topography Mission (SRTM) (<https://www2.jpl.nasa.gov/srtm>). Maps of heights, slopes, and flow accumulation models were created. The Flow accumulation model determines the natural water direction for every pixel in a DEM. Flow accumulation operation calculates the total number of pixels that will drain into certain areas (JENSON, S.K. and DOMINGUE, J.O. 1988).

Statistical analysis

Linear Least Squares Regression analysis was used to establish relationships between the values of satellite data and SOC content. A model was built separately for each band and index. Due to the limited number of soil samples available, a leave-one-out cross-validation procedure was applied (KHAN, J. *et al.* 2010; VAUDOUR, E. *et al.* 2019). The advantage of leave-one-out is that each sample participates exactly once in control from all 'n' samples within the dataset. This procedure was repeated for all n samples (GOMEZ, C. *et al.* 2012).

Prediction accuracy was evaluated by the RMSE and the R^2 values. The model with the lowest RMSE and highest R^2 values was considered as the most applicable or ideal model (JABER, S.M. *et al.* 2011). The R^2 was determined by the following classification (VAUDOUR, E. *et al.* 2019): models with $R^2 < 0.4$ show a poor or very low level of predictive ability; values of $0.5 < R^2 < 0.7$ indicate models with an average level of forecasting; models with $R^2 > 0.7$ are highly predictive.

The accuracy of the model was also determined by the classification, where RPD (re-

Table 3. Chemical shifts of atoms of the ^{13}C molecular fragments of humic acids

Chemical shift, ppm	The type of molecular fragments
0–46	C, H-substituted aliphatic fragments
46–60	Methoxy and O, N-substituted aliphatic fragments
60–110	Aliphatic fragments doubly substituted by heteroatoms (including carbohydrate) and methine carbon of ethers and esters
110–160	C, H-substituted aromatic fragments; O, N-substituted aromatic fragments
160–185	Carboxyl groups, esters, amides, and their derivatives
185–200	Quinone groups; Groups of aldehydes and ketones

sidual prediction deviation) was calculated. RPD values < 1.0 indicate a poor predictive model; 1.0 < RPD < 1.4 indicate a weak model; 1.4 < RPD < 1.8 indicate a good model that can be used for evaluation; 1.8 < RPD < 2.0 indicate a good model; 2.0 < RPD < 2.5 show a very good model and values RPD > 2.5 indicate the excellent quality of the predictive model (CHANG, C.-W. et al. 2001; VISCARRA-ROSSEL, R.A. et al. 2006b).

The statistical analysis was performed using the “caret” package in R 4.0.3 (R Development Core Team, 2015) and RStudio (version 1.3.1093) (RStudio, 2015). The IDW and ordinary kriging interpolation maps were created using standard tools in QGIS.

Results and discussion

General statistics of soil properties: mean, minimum, maximum, standard deviation (SD), coefficient of variation (CV) are shown in Table 4. The values of SOC changed in the range from 1.93 to 5.52 per cent. The depth of the humic horizon is 20 to 70 cm, mean value – 46.32 cm. Spearman’s correlation (R) between SOC content in the 0–10 cm layer and the topsoil is 0.59.

The regression analysis (Figure 2) showed that the maximum values of correlation coefficients $R = 0.78$, $R^2 = 0.61$ were detected at the invisible range B12 band (SWIR, the spatial resolution of 20 m and a spectral range of

Table 4. Statistics description of Corg in the 0–10 cm layer and the depth of the humic horizon

Parameter	Corg, %	Depth of humic horizon, cm
n = 54		
Mean	3.72	46.32
Min	1.93	20,00
Max	5.52	70,00
SD	0.88	9.49
CV, %	23.65	20.49

2,190 nm). The SWIR band of Sentinel-2A for SOC mapping shows good results in other croplands studies. Thus, in the GHOLIZADEH, A. et al. (2018) study in the Czech Republic, the authors obtained the highest correlation values (R) of B4, B5, B11, and B12 bands. The correlation of the B12 band ranged from 0.29 to 0.69, depending on the field.

The calculated spectral indices (see Table 2) showed less reliable correlation results (see Figure 2). The highest correlations were obtained using NDVI ($R = 0.68$, $R^2 = 0.46$), TVI ($R = 0.67$, $R^2 = 0.45$), and EVI ($R = 0.60$, $R^2 = 0.36$), which use bands of visible red and near-infrared range in their equations.

Attempting to diagnose the spatial distribution of the topsoil using bands and indices did not lead to reliable results. Since there is a correlation between the SOC content and the topsoil depth, the highest values in regression analysis are also shown by the B12 band ($R = 0.51$, $R^2 = 0.26$).

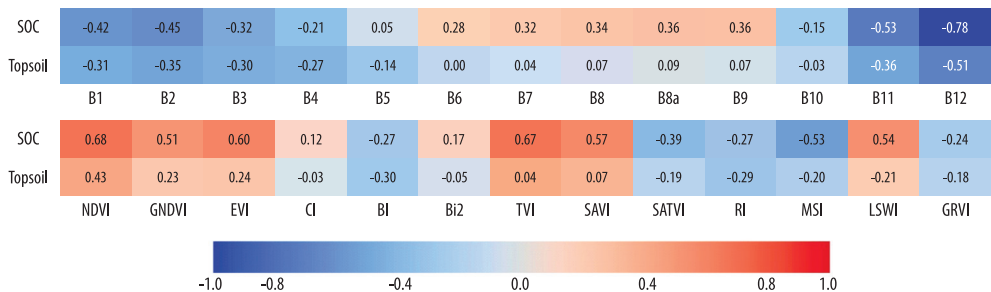


Fig. 2. The correlograms (Spearman correlation) of SOC and topsoil at bands of Sentinel-2A and calculated spectral indices. The correlation coefficients are significant at a level of 0.05.

The B12 band is the most appropriate variable for prediction SOC, according to RMSE and R^2 values (Table 5). We obtained the model with RMSE = 0.56, RPD = 1.61, and $R^2 = 0.58$. According to the classification, this model characterizes as a good model with an average prediction level (VAUDOUR, E. et al. 2019). The RPD values in GHOLIZADEH, A. et al. (2018) were 1.60–1.92 depending on the territory; similar results using Sentinel-2 data were obtained by VAUDOUR, E. et al. (2019) on the study territory of France – 1.51. The RPD results of CASTALDI, F. et al. (2019) were values 1.1–2.6 on the study areas in Germany, Belgium, and Luxembourg.

All other Sentinel bands and spectral indices are characterized by a very low level of predictive ability using a cross-validation procedure. However, the vegetation indices NDVI and TVI show values $R^2 = 0.42$. This approximation of the model to the average prediction level can be considered in future studies in similar areas and with a larger dataset.

The SOC content map based on the obtained regression equation was created using the B12 band of the satellite. Additionally, the SOC maps were created using the IDW and ordinary kriging methods to verify the spatial distribution based on the regression equation (Figure 3).

The comparative analysis of three models (regression analysis, IDW, ordinary kriging)

showed that the largest areas with the highest SOC content (4.5–5.5%) are concentrated in the western, north-western, and northern parts of the investigated field. These areas are characterized by the highest elevation elements with slopes up to 4°. Based on regression analysis, we can also observe small areas with high SOC values in the north and central parts. The areas with the smallest SOC content are located in the south-east, north-east, and central parts of the cropland. These areas are mainly located at heights between ≈ 420 and 377 m. The areas of maximum topsoil values are located at the top of the plot in the western part and also small areas in the central part. Analysis of SOC and topsoil maps revealed a spatial correlation: areas with layer thicknesses of 50 to 70 cm are equivalent to areas of SOC content 3.5–5.5 per cent (see Figure 3).

The water flows have been identified using the Flow Accumulation model (Figure 4). The main powerful flows gather throughout the site, forming the main “arteries”. The nature and direction of water flows are fully comparable with the nature of the territory relief: the main flows are concentrated in the centre of the site (direction from north-west to south-east), as well as in the northern and southern parts.

When verifying the Flow Accumulation model with field surveys, it was found that the strongest degradation processes occur in the southern part of the territory. This distribution is explained by the lowest part of the site and an increase in slope steepness (up to 6°). However, based on the analysis and maps obtained, the SOC content is not defined as homogeneous in the southern part of the area. The southern area is predominantly characterized by an average thickness surface horizon (up to 50 cm) and not high SOC content (1.5–3.5%). The study region (Trans-Ural steppe zone) is characterized by active wind erosion processes (KHAZIEV, F.KH. 1995). Thus, we can observe small plots with high SOC content (3.5–5.5%) in areas closer to and along the road due to the accumulation of soil on the leeward side of these barriers (see Figure 3). Nevertheless, this distribution can also be

Table 5. Cross-validation performance statistics

Band	RMSE	R^2	Index	RMSE	R^2
B1	0.81	0.11	NDVI	0.66	0.42
B2	0.81	0.14	GNDVI	0.77	0.21
B3	0.86	0.04	EVI	0.72	0.31
B4	0.88	0.00	CI	0.89	0.04
B5	0.89	0.22	BI	0.87	0.02
B6	0.86	0.03	BI2	0.89	0.00
B7	0.85	0.05	TVI	0.66	0.42
B8	0.85	0.05	SAVI	0.74	0.26
B8a	0.84	0.07	SATVI	0.82	0.10
B9	0.85	0.06	RI	0.87	0.02
B10	0.89	0.02	MSI	0.76	0.23
B11	0.76	0.23	LSWI	0.76	0.24
B12	0.56	0.58	GRVI	0.87	0.00

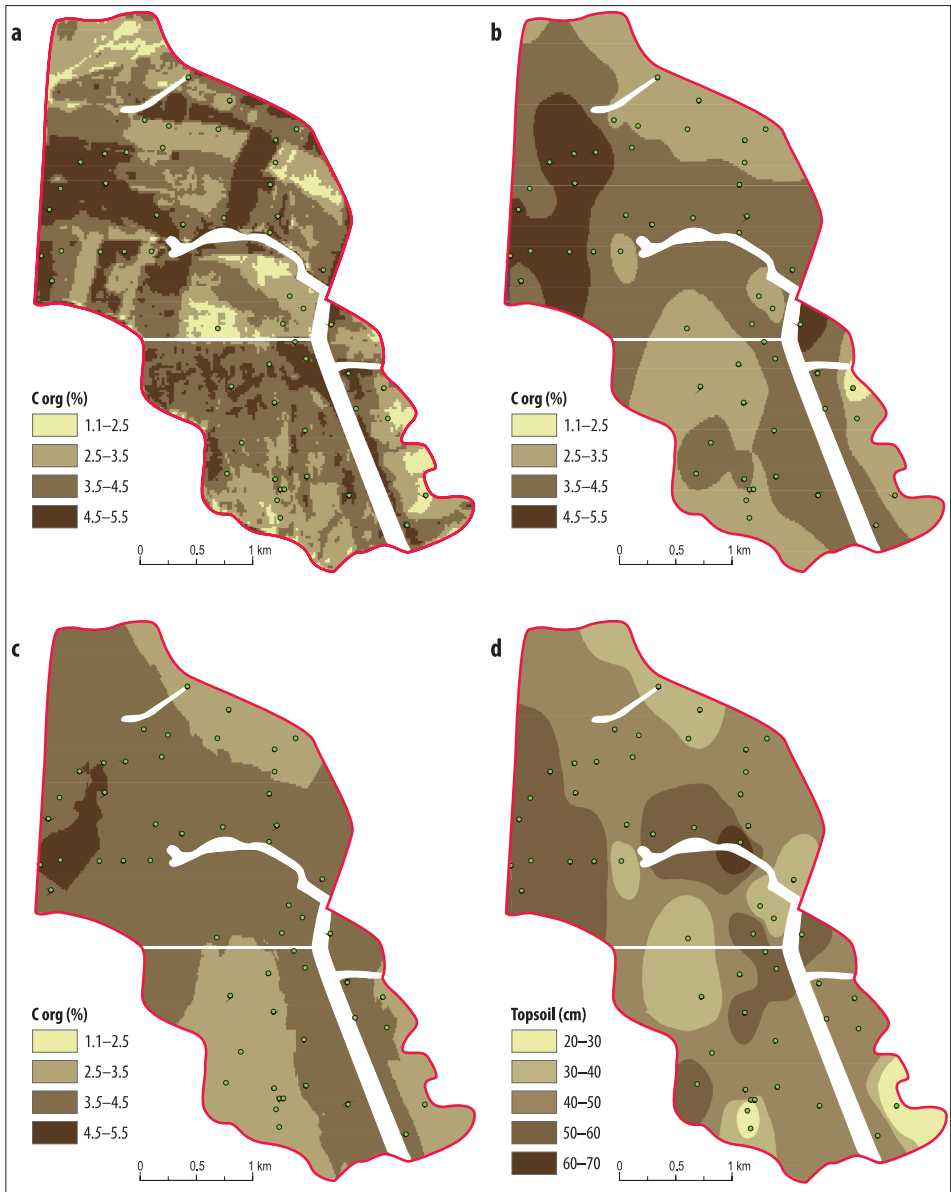


Fig. 3. The SOC content (0–10 cm) maps created using methods: a = Regression analysis based on satellite band B12; b = IDW; c = Ordinary kriging, and topsoil depth map using IDW method; d = Areas with vegetation and roads are masked by white colour.

caused by the influence of vegetation. We have masked areas with vegetation and roads, but the spatial resolution of the B12 band (20 m) can still account for this information.

The small areas of high SOC content (3.5% and more) are observed in the northern part, near the boundary of the field, which may well be consistent with the transfer of SOC

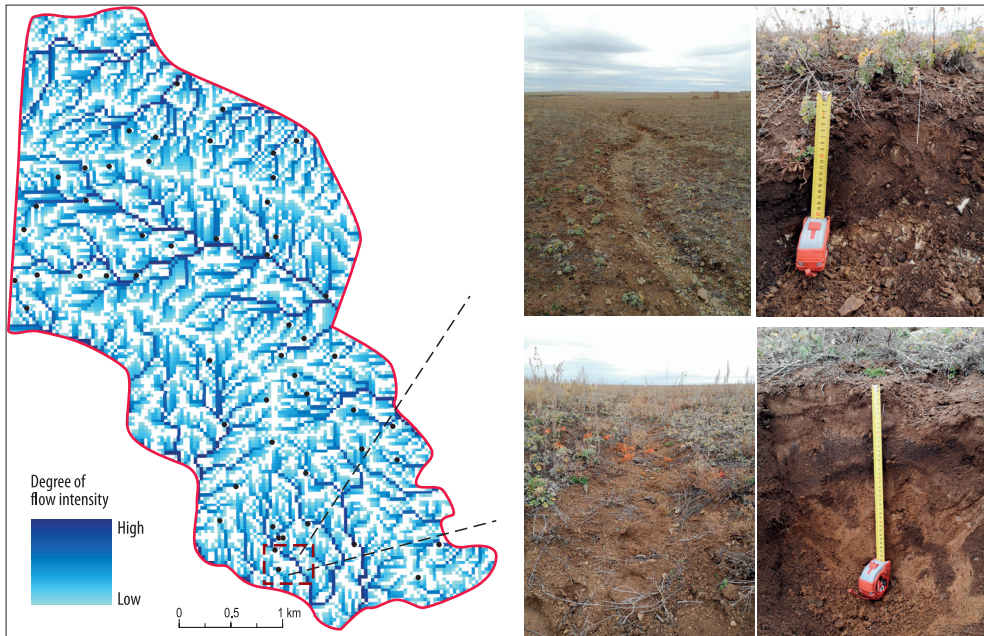


Fig. 4. Flow accumulation model map (left) and examples of rill erosion (right)

from the upper elements of the relief: this area is also characterized by lowering the relief and slopes up to 6–8°. Moreover, water erosion processes are actively occurring in this part of the field.

The lowest value of the microbiological activity of soils in the topsoil was detected in erosion sediment – 12.6 CO₂ g /100 g day⁻¹. Whereas in a non-erosive sample located a few meters from the erosion sediment, basal respiration is equal to 18.8 CO₂ g /100 g day⁻¹. The highest values are determined on the upper non-erosion elements of the terrain on average 25 CO₂ g /100 g day⁻¹ (n = 7, SD = 5). Thus, the

high microbiological activity is noted in not degradation process areas with the largest SOC content. The lower values of basal respiration are detected in an area vulnerable to erosion processes and the erosive sediment sample.

Verification of eroded and non-eroded soil samples of the south area of the field by ¹³C-NMR spectroscopy revealed the following results (Table 6). The sample No 1 (erosion sediment) is characterized by an increase in the aliphatic and oxygen-containing group compared to other samples; the ratio of AR/AL is 0.67. The erosion process led to a decrease in the aromaticity of HAs and the

Table 6. Percentage of carbon in the main structural fragments of HAs from the studied surface soil horizons*

Sample	Chemical shifts, ppm						AR	AL	AR/AL	AL h _r + AR h _r , %	C,H – AL / O,N – AL
	0–46	46–60	60–110	110–160	160–185	185–200					
1	25	6	24	30	10	5	40	60	0.67	79	0.83
2	23	6	22	33	12	4	45	55	0.82	78	0.82
3	25	6	25	27	11	6	38	62	0.61	77	0.81

*According to ¹³C-NMR data. Note: AR = Aromatic fraction; AL = Aliphatic fraction; AL h_r + AR h_r = Hydrophobicity degree in per cent; C,H – AL/O, N – AL = The degree of decomposition of organic matter.

removal of stable soil carbon. The formation of aromatic components in the soil is a long-term thermodynamic process. We assume that under conditions of water erosion, the formation of long carbon chains ($-C-C-$) and oxygen-containing fragments ($O-CH-$) occurs. This distribution also occurs in water-logging conditions (LODYGIN, E.D. et al. 2001, 2014). Water flows prevent the processes of decomposition of plant residues in the soil due to erosion processes. It leads to the acceleration of the transformation processes.

The sample No 2, which is not affected by degradation processes, has more aromatic fragments in its composition than in the samples No 1 and 3. It is distinguished by the accumulation of aromatic and carboxylic fragments; the ratio of AR/AL is 0.82. The increase of these structures in the composition of HAs is associated with the transformation of humification precursors, especially lignin-containing plant residues. In the decomposition of plant residues, up to 30 per cent of lignin enters the soil, which during transformation is included in the composition of HAs in the form of aromatic structural units and carboxylic groups.

The eroded areas are characterized by a decrease in the aromaticity of SOC. There is the removal of dark-coloured materials of SOM and fine soil particles from such areas. They are characterized by less active processes of decomposition of plant residues and microbiological processes, and thus have lower HAs values and are visualized as lighter areas according to maps constructed by regression analysis and interpolation. Such areas of the field are also well identified according to the Flow Accumulation model, which determines the rate of water flow.

The sample No 3 is represented by a little clayey top of the area from which the sheet erosion started. According to NMR spectroscopy, more aliphatic fragments of HAs are formed here – AR/AL (0.61). Clay formations have a high heat capacity and moisture retention capacity (ABU-HAMDEH, N.H. 2003; ROZHKOVA, V.A. 2006). However, such formations are often quite dense aggregates in dry places. This soil structure affects the pene-

tration and development of the root system, soil water and air movement, CO_2 emission, erosion, nutrient retention, and biological activity (CIRIC, V. et al. 2012). Clayey particles capture nutrients from the environment well, but without sufficient moistening, they become inaccessible to the plants. We assume that moisture does not accumulate here, and there is no saturation of this area because of the location on top of the studied area. Thus, there is an oppression of the soil microbiota, which affects the lower degree of humification relative to the rest of the studied areas, which is confirmed in Figure 5.

From this diagram, we can observe the following distribution: sample No 1 (erosion sediment) has a higher degree of humification than sample No 2, which is not prone to erosion. Such a change of parameters is related to the dynamic re-deposition of small particles of soil. SOM binds well with the clay due to its large specific surface area of soil aggregates. Chemically bound organomineral compounds are removed from the soil profile under the influence of water erosion and accumulate in newly formed water flow areas. Prolonged hydration of such particles favourably affects the humification processes in soil and the formation of hydrophobic macromolecules. Thus, there is a thermodynamic selection of condensed HAs macromolecules and their stabilization.

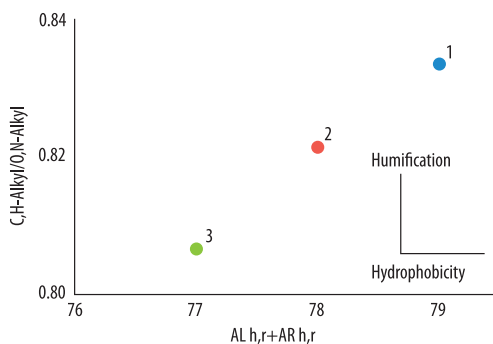


Fig. 5. The diagram of integrated indicators of the molecular composition of humic acids. $AL h,r + AR h,r$ indicates the total number of un-oxidized carbon atoms.

Verifying the obtained results of NMR spectroscopy with remote sensing and geomorphometry data, it can be stated that non-eroded areas with high SOC content are characterized by darker colouration and intensive light absorption. This is due to the high decomposition of plant residues, microbiological activity, and a high degree of aromatic humic substances in our condition. Such areas are located on top of the site, as well as areas along roads, forest plantations, and in some low-lying areas of the terrain, where valuable soil structures are transferred through water and wind erosion.

The processes of rill water erosion are shown on space images mainly as darker and spots of SOC high values due to the re-deposition of the upper fertile fractions from the top relief elements. At the same time, the darker colouration of such areas is also affected by moistening, as such deep relief elements can retain wet soil for longer, which affects the spectral reflectivity of the soil. However, rill erosion zones do not always look like darker ones, as a movement of water flows, and repositioning of fractions is affected by micro-relief. Digital models with an ultra-high spatial resolution to identify micro-reliefs are needed.

Conclusions

To date, the method of application of RS and ^{13}C -NMR spectroscopy in the investigations and digital mapping of SOC is insufficiently studied. Based on the comprehensive study of the field in long-term agricultural use with ongoing erosion processes and slopes of up to 8 degrees, it can be concluded that:

1. Sentinel-2A data can be successfully used for mapping SOC content and their uncertainty in topsoil. In our study, the highest correlation values were shown by the SWIR B12 band with a spatial resolution of 20 m and a spectral range of 2,190 nm. The developed linear regression model has an average level of prediction. The maps created allowed us to estimate the spatial distribution of SOC content on the study plot.

2. Geomorphometric analysis of the territory allowed to define more precisely the relief character and directions of water flows that determine the development of erosion processes. We can conclude that in the erosion areas, due to the active movement of soil sediments by water flows, the territory is not homogeneous in SOC content. There is an active transfer of soil fractions, which forms areas of washing away and accumulation of soil sediments. In most cases, areas along forest plantations, roads, and low elevation elements are characterized by the accumulation of SOC transported by water and wind streams from the upper parts of the relief.

3. ^{13}C -NMR analysis has shown that the non-eroded areas have a developed humic acid structure due to the complete process of decomposition of vegetation and microbial activity. Together, this has a direct impact on soil colouration and thus determines the nature of the spectral reflectance of soils. Areas vulnerable to sheet erosion are characterized by reduced aromaticity of SOC. These areas define such areas as less dark on space images. Despite less developed processes of SOC formation in areas of rill erosion, these areas are characterized by a darker colour of soils due to the re-deposition of fertile fractions and moisture accumulation.

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